



## Study on adsorption and desorption properties of the starch grafted p-tert-butyl-calix[*n*]arene for butyl Rhodamine B solution

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### ABSTRACT

The adsorbents of starch grafted p-tert-butyl-calix[4,6,8]arene-SGCn (SGC4, SGC6, SGC8) are prepared. The products are characterized by FTIR, elemental analysis, thermal gravimetric analysis and scanning electron microscope. Static adsorption behavior is studied by using SGC8 as adsorbent, butyl Rhodamine B (BRB) solution as simulation dye wastewater. The adsorption of BRB onto SGC8 fits the second order kinetic model and the apparent adsorption rate constant is  $0.002 \text{ g mg}^{-1} \text{ min}^{-1}$  at  $25^\circ\text{C}$ . The equilibrium adsorption data are interpreted using Langmuir and Freundlich models. The adsorption of BRB onto SGC8 is better represented by the Langmuir equation. The thermodynamic parameters for the adsorption reaction are calculated through van't Hoff analysis. The adsorbent may be easily regenerated by using ethanol solution as desorption agent to extract dye from SGC8. The rate of desorption of BRB is dependent on the concentration of ethanol and the temperature. SGC8 exhibits excellent adsorption and desorption properties toward dye molecule. The new-style adsorbent of SGC8 is regarded as a potential adsorbent to deal with dye or organic wastewater.

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### 1. Introduction

Dyes and pigments are widely used, mostly in leathers, textiles, paper mill, additives, foodstuff and cosmetics industry to color products. Today, they are considered as a kind of extremely important pollutant in environment due to their complex composition, high toxicity, poor degradability and great solubility in water [1,2]. Hence, the removal of dye from colored reactive dye wastewater is an important environmental issue.

The most commonly methods for the color removal include physicochemical, chemical, and biological methods, such as coagulation [3,4], precipitation [5], adsorption, filtration [6], membrane separation [7,8], chemical oxidation [9,10], ion-exchange [11,12], and aerobic and anaerobic microbial degradation [13,14]. Among these methods, liquid-phase adsorption processes are efficient for removal of dye pollutants from the industrial wastewater. More researches have been done on the adsorption behaviors of dyes in aqueous solution by adsorbents such as resins [15], chitosan [16], activated carbon [17], titanium phosphate [18], zeolites [19], iron oxides [20] and clay [21], have been reported in the literature.

Recently, much more attention has been paid to chemical separation techniques and the design and synthesis of new extraction reagents for ions and molecules. The attention results from environmental concerns and efforts to save energy and enhance recycling at the industrial level. In this respect, supramolecular chemistry has provided important avenues to prepare new type adsorbents and extraction reagents. This was achieved with the development of macrocyclic receptors, such as crown ethers [22,23], cyclodextrins [24–33] and calix[*n*]arenes [34–38]. Calix[*n*]arenes are cyclic oligomers synthesized by condensation of a p-alkylated phenol and formaldehyde. Calixarenes have a particular configuration because of its cavity. It has the flexibility to adjust the cavity dimension and the ability to form inclusion compounds with a great variety of guests, from charged molecules such as anions, metallic cations to apolar compounds. This versatility makes the calixarene family the third major class of macrocyclic binding agents after crown-ethers and cyclodextrins [39–41]. Nowadays many researchers are interested in crosslinking calixarenes with suitable crosslinkers to form an insoluble resin which has specific adsorption based on inclusion complex formation [34–38].

In this article, starch has received a considerable attention as basic matrixes for designing new adsorbent. p-Tert-butyl-calix[4,6,8]arene are grafted onto starch through epichlorohydrin to prepare adsorbents-SGCn (SGC4, SGC6, SGC8). To research the adsorption capacity and mechanism of adsorbent, butyl Rhodamine B is chosen as adsorbate (the structure is shown in Fig. 1(A)).

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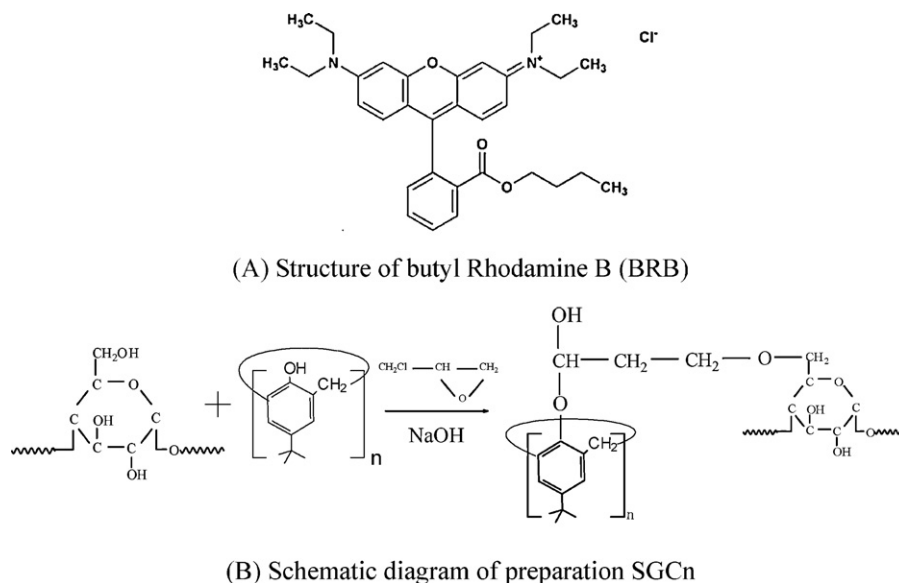


Fig. 1. (A) Structure of butyl Rhodamine B (BRB). (B) Preparation of adsorbents of starch grafted p-tert-butyl-calix[4,6,8]arene (SGCn).

Rhodamine B (RhB), Basic Violet 10, is a highly water soluble, non-volatile, basic dye of the xanthene class. It is widely used in industry as a colorant in textile, leather, jute and food industries, and is also a well-known water tracer fluorescent [42]. However, it causes phototoxic and photoallergic reactions, which have toxic, carcinogenic or mutagenic influences on living organisms [43]. Therefore, the dye-containing wastewater should be treated before discharge. The adsorption behavior of BRB onto SGC8 is studied. The cavity size effect of calixarene for the adsorption capacity is first reported. The kinetics and thermodynamic parameters for the adsorption reaction are calculated through static adsorption study. The adsorbent is regenerated by using ethanol solution as desorption agent to extract dye from SGC8. The desorption efficiency and regeneration capability are investigated and the results show that SGC8 is regarded as a potential adsorbent to deal with dye sewage.

## 2. Experimental

### 2.1. Reagent

p-Tert-butyl-calix[*n*]arene (*n* = 4, 6, 8) are synthesized according to references [44,45]. Butyl Rhodamine B (BRB) is purchased from Shanghai Chemical Reagents Company and used without purification. Starch (AR), epichlorohydrin (AR) is obtained from Sinopharm Chemical Reagent Company and used without any purification. Highly pure water is obtained from a Millipore Milli-Q UV system.

### 2.2. Instrument

FTIR spectra for the different samples are recorded on a Bruker TENSOR 27 FTIR spectrometer operated at a resolution of  $4\text{ cm}^{-1}$ . Elemental analysis for adsorbents is measured by using a SERIES II2400 (PerkinElmer) elemental analyzer. The thermal analysis of samples is characterized by thermal gravimetric analysis (TGA) on a STA409PC TGA instrument (Netzsch, Germany) with a heating rate of  $10^\circ\text{C min}^{-1}$ . The morphology and size of adsorbent are investigated by a Philips XL-30ESEM scanning electron microscope (SEM) instrument operated at an accelerating voltage of 20 kV. Fluorescence spectra are operated by F4500 (Shimadzu, Japan) using a conventional  $1\text{ cm} \times 1\text{ cm}$  quartz cell in a thermostated compartment. The excitation wavelength is 490 nm. Adsorption and

desorption experiments are completed on a laboratory shaker HZ-9211K (Jiangsu, China).

### 2.3. Preparation of adsorbents

The preparation scheme of SGC8 is shown in Fig. 1(B). The procedure is described briefly as follows: a mixture of starch (1.00 g), p-tert-butyl-calix[8]arene (4.00 g), and sodium hydroxide aqueous solution (6 mL, 20%) is stirred at  $50^\circ\text{C}$ . Then, epichlorohydrin (30 mL) is dropped into above mixture and the mixed system is stirred at  $50^\circ\text{C}$  for 24 h. The product is filtered and washed in sequence with ethanol and acetone three times. The precipitate is dried in a vacuum oven at  $50^\circ\text{C}$  for 24 h. At the end, yellow powder is obtained and characterized by FTIR, elemental analysis and SEM. SGC4 and SGC6 are synthesized by the same method.

### 2.4. Adsorption and desorption experiments

The influence of the initial solution pH on the adsorption capacity of BRB onto SGC8 is studied. Adsorption experiments are done by agitating 50 mL of dye solution at different pH ( $20\text{ mg L}^{-1}$ ) with fresh SGC8 200 mg in glass bottles using a laboratory shaker at 180 rpm and the room temperature ( $25 \pm 1^\circ\text{C}$ ). The equilibrium adsorption amount of BRB onto SGC8 at different pH is estimated by monitoring the residual BRB in the solution at the maximum emission wavelength (599 nm) by using a Fluorescence spectrophotometer. The equilibrium adsorption amount of SGC8,  $q_e$  ( $\text{mg g}^{-1}$ ), is calculated using the following relationship [46,47]:

$$q_e = \frac{(c_0 - c_e)V}{W} \quad (1)$$

where  $c_0$  is the initial concentration of BRB ( $\text{mg L}^{-1}$ ),  $c_e$  is the equilibrium concentration of dye ( $\text{mg L}^{-1}$ ),  $V$  is the volume of the solution (L), and  $W$  is the mass of the SGC8 (g).

To research adsorption kinetic behavior of BRB onto SGC8, batch adsorption experiments are done by agitating 50 mL of dye solution ( $20\text{ mg L}^{-1}$ ,  $\text{pH} = 8$ ) with fresh SGC8 200 mg in glass bottles using a laboratory shaker at 180 rpm and the room temperature ( $25 \pm 1^\circ\text{C}$ ). At different adsorption time, the dye solution is separated from the adsorbent by centrifugation at 8000 rpm for 5 min. The adsorption amount of BRB onto SGC8 at different adsorption time is estimated by monitoring the residual BRB in the solution at the maximum

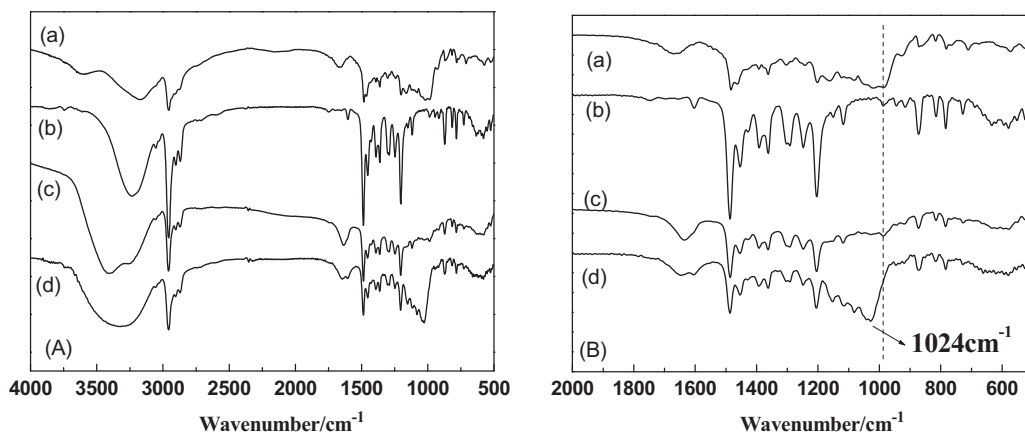


Fig. 2. FT-IR spectra of (a) starch, (b) p-tert-butyl-calix[8]arene, (c) their mixture, (d) SGC8: (A) in the range of 4000–500  $\text{cm}^{-1}$ ; (B) in the range of 2000–500  $\text{cm}^{-1}$ .

emission wavelength (599 nm) by using a fluorescence spectrophotometer.

Adsorption isotherms are obtained by using a batch equilibration technique at different temperature.

The adsorption efficiency (%), is calculated using the following relationship:

$$\eta_A = \frac{q_e}{q_m} \times 100\% \quad (2)$$

where  $q_e$  ( $\text{mg g}^{-1}$ ) is the equilibrium adsorption capacity of SGC8 and  $q_m$  ( $\text{mg g}^{-1}$ ) is the saturation adsorption capacity.

In desorption experiment, 2 g fresh SGC8 is put in 1 L 80  $\text{mg L}^{-1}$  BRB aqueous solution. The solution above is shaken by laboratory shaker (180 rpm) 5 h at 25  $^{\circ}\text{C}$  to make an equilibrium absorption of BRB onto SGC8, which is filtrated and dried at room temperature for further use. The adsorption capacity is about 12  $\text{mg g}^{-1}$ . Then, 0.1 g above SGC8 is added into conical flasks. The different concentration of ethanol solution is used as desorption agent and the volume is 20 mL. At desorption equilibrium, the concentration of BRB in desorption agent phase is measured by fluorescence spectra as previously described.

The desorption efficiency (%), is calculated using the following relationship:

$$\eta_D = \frac{c_{e,D}V}{q_m W} \times 100\% \quad (3)$$

where  $c_{e,D}$  ( $\text{mg L}^{-1}$ ) is the desorption equilibrium concentration of dye,  $V$  (L) is the volume of the desorption agent solution,  $q_m$  ( $\text{mg g}^{-1}$ ) is the saturation adsorption capacity and  $W$  (g) is the mass of the SGC8.

In a typical regeneration experiment, the 80% ethanol solution is used as desorption agent and the volume is 20 mL. 0.1 g SGC8 covered with BRB is dipped in turn in fresh desorption agent for 10 $\times$ . The adsorption and desorption efficiency is calculated by using Eqs. (2) and (3).

### 3. Results and discussion

#### 3.1. Characterization of adsorbents

The FTIR spectra of starch, p-tert-butyl calix[8]arene, their mixture and SGC8 are taken by FTIR spectrometer with a KBr film and shown in Fig. 2. Fig. 2(a) shows the FTIR spectrum of starch. The band observed at 3460  $\text{cm}^{-1}$  is assigned to –OH stretching vibration. For p-tert-butyl calix[8]arene, the characteristic vibration peaks of –OH at 3244  $\text{cm}^{-1}$ , –C(CH<sub>3</sub>)<sub>3</sub> stretching vibration at 2957  $\text{cm}^{-1}$ , phenyl plane bending vibrations at 1608, 1485  $\text{cm}^{-1}$

are clearly observed in Fig. 2(b). All above absorption peaks can be found in Fig. 2(c). However, in Fig. 2(d), the band observed at 2957  $\text{cm}^{-1}$  is assigned to the p-tert-butyl stretching vibration, which indicates powerful that p-tert-butyl-calix[8]arene is immobilized on starch. Furthermore, compared with the mechanical mixture (shown in Fig. 2B(c)), the band observed at 1024  $\text{cm}^{-1}$  is assigned to stretching vibration of C–O–C (shown in Fig. 2B(d)), which provides a substantial evidence of grafting of p-tert-butyl calix[8]arene onto the starch. The FT-IR spectra of SCG4 and SCG6 are shown in Figs. S1 and S2.

The result of elemental analysis further confirmed the immobilized p-tert-butyl-calix[8]arene on starch. Elemental analysis for p-tert-butyl-calix[8]arene (molecular formula:  $(\text{C}_{11}\text{H}_{14}\text{O})_8$ ): C, 81.37%, H, 9.11%. The ratio of C:H is very close to theoretic value of p-tert-butyl-calix[8]arene (81.48:8.64). Elemental analysis for starch (molecular formula  $(\text{C}_6\text{H}_{10}\text{O}_5)_n$ ): C, 41.65%, H, 6.46%. Elemental analysis for SGC8: C, 44.59%, H, 6.89%. The contents of carbon and hydrogen increase slightly, which indicates that the p-tert-butyl-calix[8]arene is immobilized on starch. Elemental analysis for SGC4 and SGC6 is listed in Table S1. The contents of carbon and hydrogen increase with the number of p-tert-butyl-phenol.

As shown in Fig. 3, thermal gravimetric analysis (TGA) is used to analyze the thermostability of SGC. For starch, two main steps shown in Fig. 3(a) are observed. The first step ranging from 45 to 139  $^{\circ}\text{C}$  is due to the loss of moisture. The second ranging from 278 to 335  $^{\circ}\text{C}$  is assigned to the breakage of starch backbone. Two main steps on thermal degradation curves for p-tert-butyl-calix[8]arene are shown in Fig. 3(b). The first step ranging from 366 to 410  $^{\circ}\text{C}$  is assigned to the loss of the functional group of p-tert-butyl in

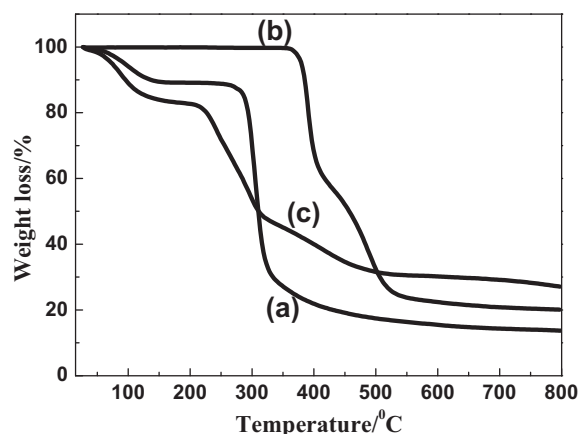


Fig. 3. TGA curves (a) starch, (b) p-tert-butyl calix[8]arene, (c) SGC8.

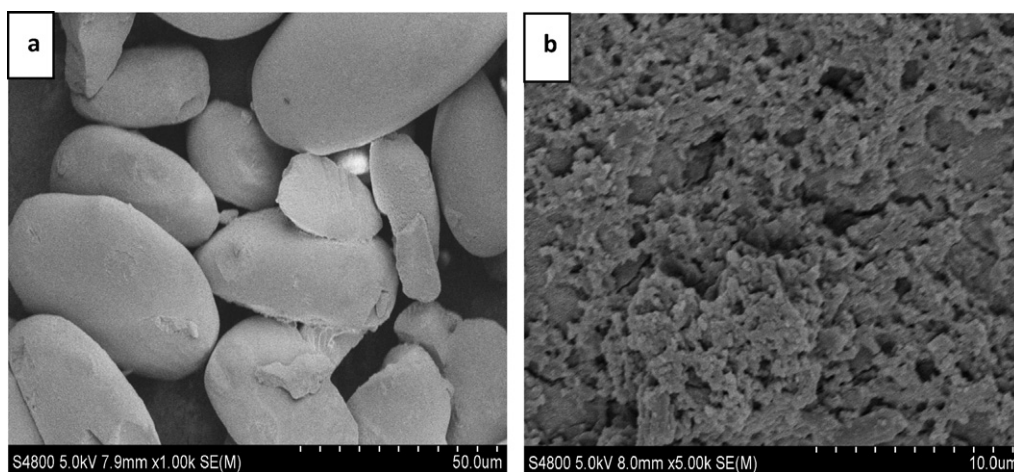


Fig. 4. SEM micrograph of (a) starch, (b) SGC8.

calixarene. The weight loss is about 38% which is similar to the theoretic value. The second step ranging from 410 to 524 °C is attributed to the breakage of the ring of the calix[8]arene. The TGA curve of SGC8 is shown in Fig. 3(c). The thermal degradation for SGC is more complicated and three main steps are observed clearly. The first and second steps ranging from 40 to 125 °C and 215–321 °C are similar to that observed for thermal degradation of starch shown as in Fig. 3(a). The degradation ranging from 321 to 530 °C may be assigned to the degradation of calix[8]arene grating onto the polymer. Thermal degradation curves for SGC4 and SGC6 are presented in Fig. S3. Compared with SGC8, the shape of the thermal degradation curves is quite similar. But the weight loss increases in the order of the size of the calixarene ring, i.e., SGC4 < SGC6 < SGC8.

Fig. 4 shows the SEM micrograph of starch (a) and SGC8 (b). It is clear that the morphology of starch changes a lot after *p*-tert-butyl-calix[8]arene immobilized on the surface of carrier, which provides an assistive technology support of grafting of *p*-tert-butyl calix[8]arene onto starch. The morphology of SGC4 (a) and SGC6 is similar to that of SGC8 and the SEM micrograph of SGC4 (a) and SGC6 (b) are shown in Fig. S4.

### 3.2. Adsorption behavior of BRB onto SGCn

#### 3.2.1. Effect of pH

The value of pH is known as one of the most important factors affecting adsorption capacity of adsorbents. Fig. 5 shows the influence of the initial solution pH on the adsorption capacity of BRB

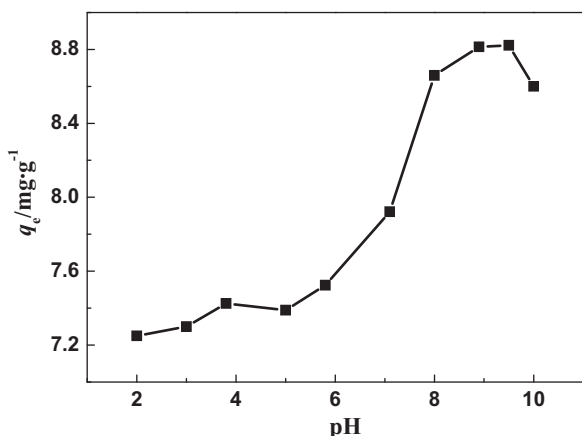


Fig. 5. At 25 °C, adsorption capacity of BRB onto SGC8 at different pH.

onto SGC8. From Fig. 5, the adsorption capacity of BRB onto SGC8 increases with the pH while the value of pH changes from 2 to 9. There is slight increase in the adsorption capacity at low acidity. On the contrary, the significant change in the adsorption capacity at the high value of pH is shown in Fig. 5. The largest adsorption capacity is 9.81  $\text{mg} \cdot \text{g}^{-1}$  at pH = 9.0. Several investigations report that adsorption quantity of basic dye usually increases as the pH is increased [48,49]. Basically, at lower value of pH, positively charged surface sites on the adsorbent do not favor the adsorption of dye cations due to the electrostatic repulsion. At high pH,  $\text{OH}^-$  on the surface of adsorbent will favor the adsorption of cationic dye molecules. However, in the application of dye wastewater treatment, the value of pH of dye solution is always regulated to neutral. Therefore, from the results of experiment and practical application, pH = 8.0, weak alkalinity, is chosen as the appropriate value of the initial solution of BRB.

#### 3.2.2. Kinetic behavior of adsorption

From batch adsorption experiments, the adsorption capacity of BRB onto starch, SGC4, SGC6 and SGC8 at different adsorption time are shown in Fig. 6. It clearly shows that the adsorption capacity of starch, SGC4 and SGC6 is increased slowly with the time during the first 60 min and adsorption equilibrium reach after 100 min. However, when SGC8 is used as adsorbent, the adsorption capacity

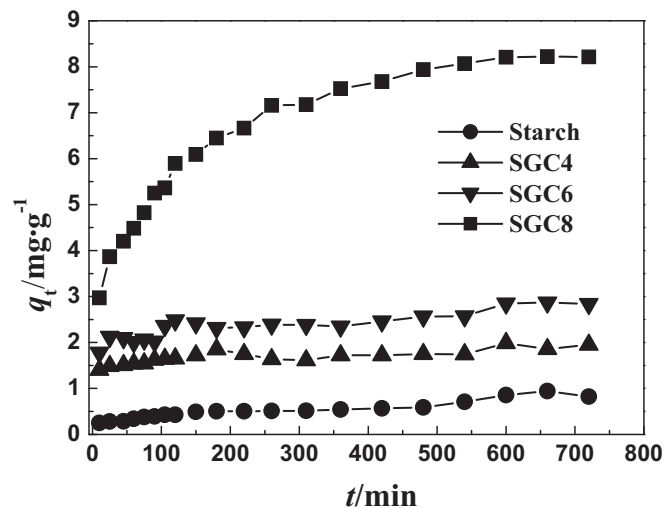


Fig. 6. The relationship between  $q_t$  and  $t$  of different adsorbents.



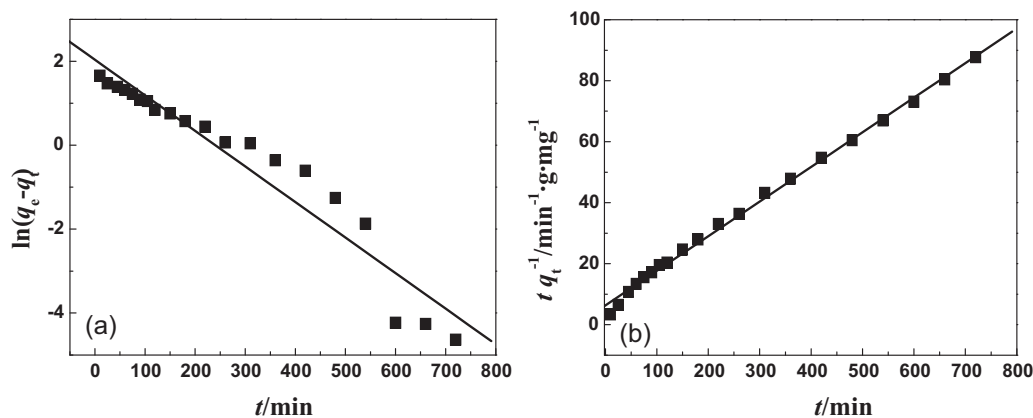


Fig. 7. (a) The relationship between  $\ln(q_e - q_t)$  versus  $t$ . (b) The relationship between  $(t/q_t)$  versus  $t$ .

is increased rapidly with the time and an adsorption equilibrium reaches after 600 min. It is evident that the equilibrium adsorption capacity of SGC8 is further larger than that of starch, SGC4 and SGC6. So, SGC8 is chosen the main object of study.

Many kinetic models are developed in order to find intrinsic kinetic adsorption constants. Traditionally, the kinetics of metal ions adsorption is described following the expressions originally given by Lagergren which are special cases for the general Lagergren rate equation [50]. A simple kinetic analysis of adsorption is the pseudo-first-order equation in the form:

$$\ln(q_e - q_t) = -k_1 t + \ln q_e \quad (4)$$

where  $k_1$  ( $\text{min}^{-1}$ ) is adsorption rate constant,  $q_e$  and  $q_t$  ( $\text{mg g}^{-1}$ ) adsorption capacity of dye adsorbed at equilibrium and at any time, respectively.

A pseudo-second-order equation based on adsorption equilibrium capacity may be expressed as follows [51,52]:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \quad (5)$$

where  $k_2$  ( $\text{g mg}^{-1} \text{min}^{-1}$ ) is adsorption rate constant of a pseudo-second-order equation.

The fitting validity of these models is traditionally checked by each linear plot of  $\ln(q_e - q_t)$  versus  $t$ , and  $(t/q_t)$  versus  $t$ , respectively. Fig. 7(a) and (b) show the fitting plots of two kinetic models. From correlation coefficients, the value of  $R^2$  in the second-order kinetic model (0.997) is higher than the first-order kinetic model value (0.927), which indicates that the adsorption of BRB onto SGC8 fits a pseudo-second-order reaction. From the slope and intercept of

the straight line (Fig. 7(b)), the pseudo-second-order rate constant  $k_2$  is evaluated as  $0.002 \text{ g mg}^{-1} \text{ min}^{-1}$ .

### 3.2.3. Adsorption isotherms

The adsorption isotherms of BRB onto starch, SGC4, SGC6 and SGC8 at  $25^\circ\text{C}$  are shown in Fig. 8(a). Comparing adsorption capacity of four adsorbents at  $25^\circ\text{C}$ , the saturation adsorption capacity increases in the order of starch, SGC4, SGC6 and SGC8. In three starch-calix[8]arene polymer, the saturation adsorption capacity of SGC8 is larger than that of SGC4 or SGC6, which is attributed to the larger formation constant between calix[8]arene and BRB (The formation constants between calix[4,6,8]arene and BRB are calculated by fluorescence spectroscopy). Therefore, the larger formation constant between calixarene and guest molecule is, the larger adsorption capacity of the adsorbent is.

The adsorption isotherms of BRB onto SGC8 at different temperature are shown in Fig. 8(b). The Langmuir sorption isotherm is the most widely used for the adsorption of a pollutant from a liquid solution assuming that the adsorption takes place at specific homogeneous sites within the adsorbent [53,54]. Assuming the adsorption coincided with Langmuir's model, the Langmuir adsorption isotherm is expressed by the following equation [53,54]:

$$\theta = \frac{q_e}{q_m} = \frac{K_L c_e}{1 + K_L c_e} \quad (6)$$

where  $\theta$  is coverage,  $c_e$  ( $\text{mg L}^{-1}$ ) and  $q_e$  ( $\text{mg g}^{-1}$ ) are the liquid phase concentration and solid phase concentration of sorbent at equilibrium, respectively;  $q_m$  ( $\text{mg g}^{-1}$ ) is the saturation adsorption capacity and  $K_L$  ( $\text{L mg}^{-1}$ ) is the Langmuir isotherm constant.

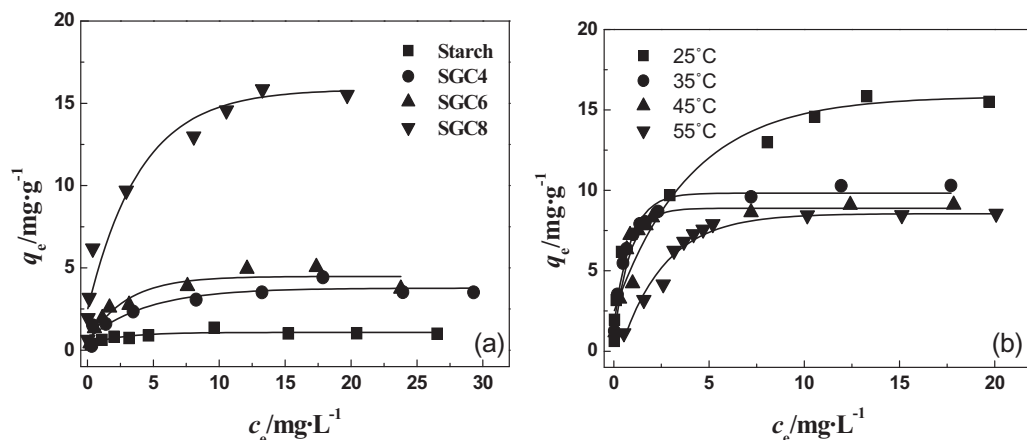


Fig. 8. (a) At  $25^\circ\text{C}$ , adsorption isotherm for BRB onto different adsorbents. (b) Adsorption isotherm for BRB onto SGC8 at different temperature.

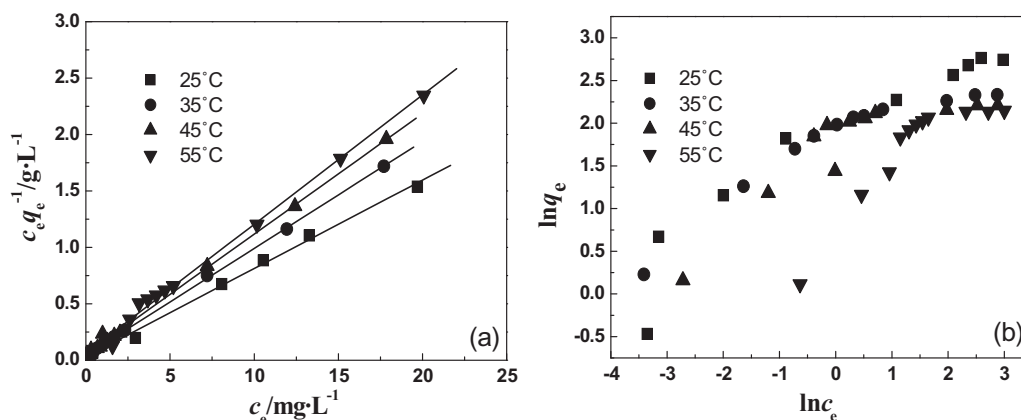


Fig. 9. (a) The relationship between  $c_e/q_e$  and  $c_e$  at different temperature. (b) The relationship between  $\ln q_e$  and  $\ln c_e$  at different temperature.

The linear equation of Langmuir is represented as follows:

$$\frac{c_e}{q_e} = \frac{c_e}{q_m} + \frac{1}{K_L q_m} \quad (7)$$

Fig. 9(a) shows the relationships between  $c_e/q_e$  and  $c_e$  at different temperature. A well behaved linear relationship implies that adsorption of BRB onto SGC8 obeys Langmuir's model. According to the slope and the intercept of the straight lines,  $q_m$  and  $K_L$  both can be evaluated. Table 1 lists the values of  $q_m$ ,  $K_L$  and  $R^2$  ( $R$ , the correlation coefficient) at different temperature for Langmuir isotherm. From the value of  $R^2$ , the adsorption system under the concentration range studied fits Langmuir isotherm well. From Table 1, both  $q_m$  and  $K_L$  are dependent on the temperature. With the temperature increasing, the values of  $q_m$  and  $K_L$  decrease gradually, which indicates that the higher temperature, the more disadvantageous for the adsorption of BRB onto SGC8.

It is well known that the Freundlich isotherm is the earliest known relationship describing the adsorption equation [55]. The application of the Freundlich equation suggests that sorptional energy exponentially decreases on completion of the adsorption centers of a sorbent. This isotherm is an empirical equation employed to describe heterogeneous systems and is expressed by the following equation:

$$q_e = K_F c_e^{1/n} \quad (8)$$

where  $q_e$  is the equilibrium dye concentration onto adsorbent ( $\text{mg g}^{-1}$ ),  $c_e$  is the equilibrium dye concentration in solution ( $\text{mg L}^{-1}$ ),  $K_F$  is Freundlich constant, and  $1/n$  is the heterogeneity factor. The capacity constant  $K_F$  and the affinity constant  $n$  are empirical constants dependent on several environmental factors. A linear form of the Freundlich expression can be obtained by taking logarithms of Eq. (9):

$$\ln q_e = \ln K_F + \frac{1}{n} \ln c_e \quad (9)$$

The plot of  $\ln q_e$  versus  $\ln c_e$  is shown in Fig. 9(b). Table 1 lists the values of  $K_F$ ,  $n$  and  $R^2$  at different temperature for Freundlich isotherm.

**Table 1**  
At different temperature adsorption constants and the correlation coefficients for Langmuir and Freundlich isotherm.

| T/K | Langmuir isotherm      |                        |       | Freundlich isotherm |      |       |
|-----|------------------------|------------------------|-------|---------------------|------|-------|
|     | $K_L/\text{L mg}^{-1}$ | $q_m/\text{mg g}^{-1}$ | $R^2$ | $K_F$               | $n$  | $R^2$ |
| 298 | 2.66                   | 12.76                  | 0.997 | 5.53                | 2.39 | 0.888 |
| 308 | 2.25                   | 10.54                  | 0.998 | 5.81                | 3.22 | 0.846 |
| 318 | 2.11                   | 9.37                   | 0.998 | 5.10                | 3.13 | 0.721 |
| 328 | 1.93                   | 8.73                   | 0.994 | 2.56                | 1.87 | 0.785 |

From the value of  $R^2$ , the linear relationship does not fit well for the adsorption system under the concentration range studied. Apparently, this plot demonstrates that equilibrium adsorption data of BRB is not excellent for the Freundlich isotherm. It is evident that the Langmuir model is better than the Freundlich model in this case. The Langmuir equation is applicable to homogeneous adsorption, where the adsorption of each molecule onto the surface has equal adsorption activation energy. This suggests that some homogeneity in the surface of sorbent will play a key role in dye adsorption and that the homogeneous adsorption model will be better for isotherm simulation. The value of the monolayer saturation capacity indicates that SGC8 exhibits interesting adsorption properties toward dye molecule and SGC8 will be a potential adsorbent for dealing with dye or organic wastewater.

### 3.2.4. Thermodynamics parameters of adsorption reaction

To elucidate the thermodynamic origins of the adsorption reaction of BRB onto SGC8, the thermodynamic parameter for the adsorption reaction is determined through van't Hoff analysis. The relationship of the Langmuir isotherm constant  $K_L$  with temperature  $T$  can be described by van't Hoff equation shown as follows:

$$\ln K_L = -\frac{\Delta H^\theta}{RT} + \frac{\Delta S^\theta}{R} \quad (10)$$

where  $R$  is the gas constant,  $\Delta H^\theta$  is the enthalpy change and  $\Delta S^\theta$  is the entropy change. The  $K_L$  ( $\text{L g}^{-1}$ ) has to be converted to  $K_L'$  ( $\text{L mol}^{-1}$ ) by using the molecular weight of BRB before calculating  $\Delta H^\theta$  and  $\Delta S^\theta$  [56].  $\Delta H^\theta$  ( $\text{kJ mol}^{-1}$ ) and  $\Delta S^\theta$  ( $\text{J mol}^{-1} \text{K}^{-1}$ ) are calculated from the slope and intercept of linear plot of  $\ln K_L'$  versus  $T^{-1}$  (the plot is shown in Fig. 10).

The Gibbs free energy change,  $\Delta G^\theta$ , is estimated from the following relationship:

$$\Delta G^\theta = \Delta H^\theta - T\Delta S^\theta \quad (11)$$

The thermodynamic parameter for the adsorption reaction is determined through van't Hoff analysis. The good linear relationship between  $\ln K_L'$  and  $1/T$  is obtained (Fig. 10), and therefore the value of the molar enthalpy, the entropy change and the Gibbs free enthalpy change is calculated  $-8.41 \text{ kJ mol}^{-1}$ ,  $-7.25 \text{ J mol}^{-1} \text{K}^{-1}$  and  $-6.25 \text{ kJ mol}^{-1}$ , respectively. The adsorption reaction of BRB onto SGC8 is driven by the favorable enthalpic change, accompanying negative entropy change. The negative value of molar enthalpy change indicates that the adsorption reaction is exothermic reaction, so higher temperature is disadvantageous to the adsorption reaction. The negative value of entropy change is due to the decrease of the freedom of dye molecule after the adsorption of BRB onto SGC8.

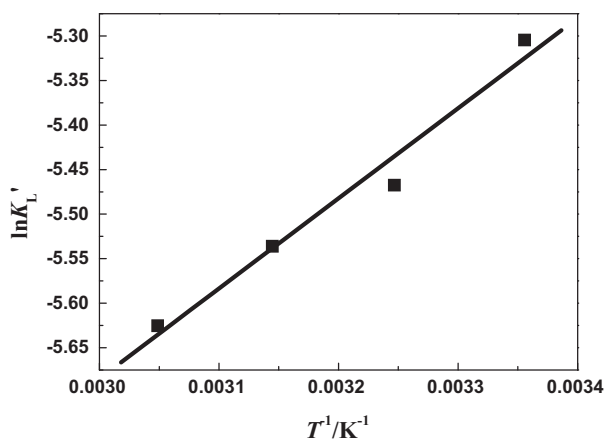


Fig. 10. Relationship between  $\ln K'_L$  and  $T^{-1}$ .

Table 2

The desorption efficiency (%) with different concentration of ethanol solution as desorption agent.

|                         | Ethanol/% |      |      |      |      |
|-------------------------|-----------|------|------|------|------|
|                         | 0         | 20   | 40   | 60   | 80   |
| Desorption efficiency/% | 1.97      | 23.6 | 54.1 | 71.4 | 99.7 |

### 3.3. Desorption of BRB from SGC8

#### 3.3.1. Effect of concentration of desorption agent

In desorption experiment, the different concentration of ethanol solution is used as desorption agent. The desorption equilibrium time is about 9 h. The rate of desorption (%), is calculated by Eq. (3) and the result is shown in Table 2. It indicates that the rate of desorption depends on the concentration of ethanol. When the concentration of ethanol is 80%, the desorption efficiency is close to 100%, which reveals that higher concentration of ethanol is good desorption agent.

#### 3.3.2. Temperature dependence of the desorption

Fig. 11 shows the rate of desorption at different temperature with 40% ethanol solution. The amount of BRB desorbed from SGC8 depends on the temperature. It is clear that the desorption efficiency of BRB from SGC8 increases with temperature, which demonstrates that higher temperature is advantageous to the desorption reaction.

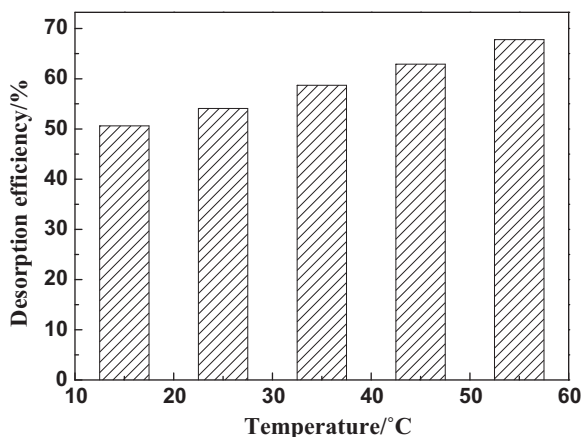


Fig. 11. The plot of the desorption efficiency at different temperature using 40% ethanol solution as desorption agent.

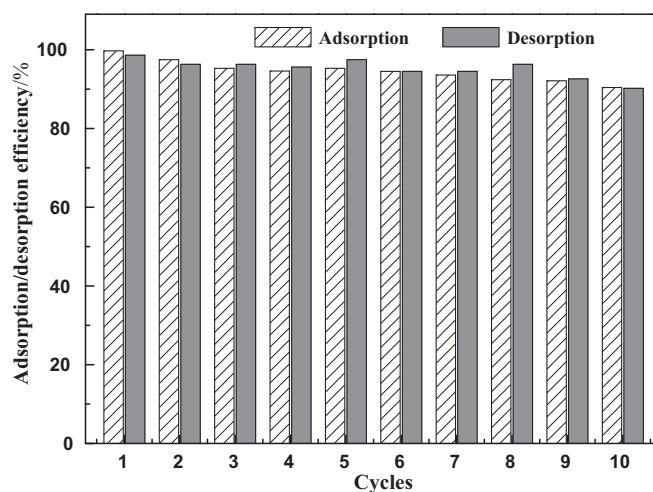


Fig. 12. The adsorption/desorption efficiency of cycles for BRB onto SGC8 at 25 °C.

#### 3.3.3. Regeneration of the adsorbent

To the practical application of the adsorption process using SGC8, it is necessary to study the regeneration of the adsorbent. 80% ethanol solution is used as desorption agent for SGC8 regeneration. The adsorbent is used 10 times repeatedly and the rates of adsorption and desorption are shown in Fig. 12. With the increasing of the repeated times, the rates of adsorption and desorption are decreased slightly. It is found that SGC8 can be used repeatedly without significant decline of its adsorption capabilities during adsorption–desorption cycles. However, using ethanol solution as desorption agent will cause secondary pollution and the cost is too expensive. Therefore, if the desorption agent is improved, SGC8 can be regarded as a potential adsorbent in the application of sewage treatment.

## 4. Conclusions

The adsorbents of starch grafted p-tert-butyl-calix[4,6,8]arene-SGCn (SGC4, SGC6, SGC8) are synthesized successfully. The adsorption of BRB onto SGC8 fits the second order kinetic model. The order of the saturation adsorption capacity at 25 °C is consistent with the size of the calixarene ring, i.e., SGC4 < SGC6 < SGC8. The adsorption behavior of BRB onto SGC8 obeys Langmuir's model. The thermodynamic parameters indicate that the adsorption reaction of BRB onto SGC8 is driven by the favorable enthalpic change, accompanying negative entropy change. The desorption efficiency of BRB is dependent on the concentration of ethanol and the temperature. The adsorbent may be easily regenerated by using 80% ethanol solution as desorption agent to extract dye from SGC8. The adsorbent of SGC8 can be regarded as a potential adsorbent in the application of dye or organic wastewater treatment.

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## Appendix A. Supplementary data

Supplementary data associated with this article can be found, in the online version, at doi:10.1016/j.jhazmat.2010.09.107.

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